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Chlorinated Byproducts of Neonicotinoids and their Metabolites:

An Unrecognized Human Exposure Potential?

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1 ABSTRACT

2 We recently reported initial discovery of neonicotinoid pesticides in drinking water and potential 3 for transformation through chlorination and alkaline hydrolysis during water treatment. The <mark>4</mark> objectives of this research were to determine: (1) if neonicotinoid metabolites are relevant to drinking water exposure, and (2) the products formed from chlorination of neonicotinoids and their 5 6 metabolites. Desnitro-imidacloprid and imidacloprid urea, two known metabolites of imidacloprid, 7 are documented for the first time in drinking water. Desnitro-imidacloprid was present above the 8 lower level of detection (0.03 ng/L) in 67% of samples (6/9) from drinking water systems but 9 detectable in all samples (up to 0.6 ng/L). Although concentrations of desnitro-imidacloprid were lower than concentrations of parent neonicotinoids, desnitro-imidacloprid exhibits significantly 10 more mammalian toxicity than imidacloprid. Using LC-HR-ToF-MS/MS analysis of laboratory 11 experiments, we propose structures for novel transformation products resulting from the 12 chlorination of clothianidin, imidacloprid, desnitro-imidacloprid, imidacloprid-urea, and 13 hydrolysis products of thiamethoxam. Formation of chlorinated neonicotinoid byproducts occurs 14 at timescales relevant to water treatment/distribution for the imidacloprid metabolites ($t_{1/2}$ =2.4min-15 1.0h) and thiamethoxam hydrolysis products (4.8h). Imidacloprid metabolites in finished drinking 16 17 water and potential formation of novel disinfection byproducts during treatment/distribution are 18 relevant to evaluating the exposure and potential impacts of neonicotinoids on human health.

19 **TOC Art**:



21 INTRODUCTION

22 Neonicotinoids are the most widely used insecticides in the world.¹ Neonicotinoids are 23 systemic, insect-targeting neurotoxins that have gained popularity due to their broad spectrum of control, high potency, and insect selectivity.²⁻⁴ This insecticide class enjoys a wide range of both 24 urban and agricultural uses, with a majority (~80% annually) of treated seeds planted in the United 25 States coated with neonicotinoids.^{1,5} Due to chemical properties (polarity, solubility) and heavy 26 usage, neonicotinoids are commonly measured in surface waters across North America⁶⁻¹⁰ with 27 reported concentrations^{7,11–14} up to 6900 ng/L. Neonicotinoid metabolites, such as desnitro-28 29 imidacloprid and imidacloprid-urea, are formed via microbial degradation, as well as some abiotic processes (e.g., photolysis, hydrolysis).^{2,3,5,15–22} As a result, these metabolites may also be present 30 in surface waters used for drinking water. 31

32 Neonicotinoids exploit specific differences between nicotinic acetylcholine receptors (nAChR) in vertebrates and invertebrates to impart insect selectivity.^{2,23} Neonicotinoids share important 33 functional groups (nitroimines, cyanoimines, or nitromethylenes) to influence electrostatic binding 34 potential; the negative polarity^{24,25} on the neonicotinoid is rejected by the mammalian nAChR and 35 readily accepted by the insect nAChR.² Although selective toxicity improves safety for non-target 36 vertebrate organisms, the effects of chronic exposure of humans to neonicotinoids remain 37 unknown.^{26,27} Furthermore, toxicological profiles of neonicotinoid transformation products 38 39 formed via degradation processes may be different from that of the parent compounds, particularly 40 when the nitro- or cyano-groups are removed. For example, two known metabolites of imidacloprid and thiacloprid—desnitro-imidacloprid and descyano-thiacloprid—are respectively 41 317 and 195 times more toxic to mammals (based on IC_{50}) than their corresponding parent 42 compounds.³ Understanding the identity, fate, and bioactivity of transformation products 43

generated in natural and engineered systems is critical to understanding the full impacts ofneonicotinoids on ecosystems and human health.

We recently reported²⁸ the first measurement of neonicotinoids in finished drinking water and 46 47 demonstrated that select neonicotinoids can be transformed at elevated pH (thiamethoxam) or during chlorination (clothianidin, imidacloprid) over timescales relevant to water treatment and 48 49 distribution. There is increasing concern about anthropogenic compounds acting as disinfection byproduct (DBP) precursors during disinfection²⁹ and the potential for these next-generation DBPs 50 to exhibit retained or even enhanced bioactivity³⁰ (i.e., carcinogenic and/or genotoxic³¹). 51 52 Objectives of this research were to determine: (1) if neonicotinoid metabolites are relevant to drinking water exposure, and (2) the products formed from chlorination of neonicotinoids and their 53 54 metabolites that may be generated during drinking water treatment.

55

56 MATERIALS and METHODS

Drinking water samples. Raw and treated (entering and exiting treatment plant, respectively) 57 drinking water samples were collected from the University of Iowa (UI) and Iowa City (IC) 58 59 drinking water treatment plants (Iowa City, IA, USA). The treatment trains are detailed in the SI (Scheme S.1). The main similarities are both systems use lime softening at elevated pH (>10.3) 60 and free chlorine disinfection, the main differences are that UI uses direct surface water and 61 conventional coagulation/flocculation/sand-filtration with mixed powdered activated carbon 62 during high dissolved organic matter (DOM) conditions (to control DBP formation), whereas IC 63 uses an alluvial well-field with granulated activated carbon (GAC) filter-beds. Tap samples were 64 collected from two buildings on the UI campus and three residences serviced by the IC plant 65

located throughout the city. The limited number of samples was intended to establish the presence 66 and relevance of neonicotinoid metabolites in drinking water, but was not intended to be fully 67 spatially / temporally representative, nor collected in a Lagrangian manner (i.e., transport time-68 adjusted). Samples were collected during the summer months, when neonicotinoid concentrations 69 are highest.^{6,28} UI and IC drinking water samples were analyzed for clothianidin, imidacloprid, 70 71 thiamethoxam, desnitro-imidacloprid, and imidacloprid-urea. Methods for sample collection and analysis, as well as background information for both treatment and distribution systems, are 72 described previously.²⁸ Analytical details, lower limits of detection (LLD), and field blank data 73 74 are provided in the SI.

Hydrolysis, chlorination, and transformation product analysis. Fate during unit processes 75 76 (lime softening, disinfection, and sequential lime softening and disinfection) was simulated in laboratory batch systems (described fully in the SI) using pH adjustment and free chlorine addition 77 with neonicotinoid concentrations measured by liquid chromatography with diode array detector 78 79 (LC-DAD). Experiments used free chlorine (HOCl) in a closed reactor containing 5 mM phosphate buffer (pH 7); a range of neonicotinoid (1–50 µM) and HOCl (1–50 mg/L) concentrations were 80 81 tested (described in Figures S.2, S.3). Chlorination of thiamethoxam hydrolysis products occurred following initial hydrolysis at elevated pH with no chlorine (details in SI). Samples were monitored 82 for 24-72 hours via LC-DAD, and then brought to the High-Resolution Mass Spectrometry Facility 83 (HRMSF) at the University of Iowa for exact mass identification and MS/MS fragment analysis 84 via LC-HR-ToF-MS/MS (Figures S6-S40). The Schymanski framework³² was used for 85 86 communicating confidence in identifying newly discovered small molecules (Table 1). Stability of chlorinated products (DN-IMI 245 chosen as representative example) was examined by adding 87

freshly-prepared sulfite (50 μM in the reactor) and observing back-transformation via LC-MS.
Experimental details and analytical methods are provided in the SI.

90 **RESULTS AND DISCUSSION**

91 Occurrence of neonicotinoids and their metabolites in drinking water samples. Desnitroimidacloprid was present above the Lower Level of Detection (LLD)³³ in 67% (6/9) of samples 92 93 (raw, treated, and tap water) collected from UI (4/4) and IC (2/5) drinking water systems (Figure 94 1) but was detectable above the instrument signal-to-noise in all samples analyzed, representing 95 the first known documentation of neonicotinoid metabolites in drinking water. The concentration of desnitro-imidacloprid ranged from <0.03-0.60 ng/L for all water samples. The desnitro-96 97 imidacloprid tap water concentrations ranged from 0.03-0.06 ng/L at UI and <0.03 ng/L for IC. 98 Imidacloprid-urea was also present above the LLD in 56% (5/9) of all samples analyzed (4/4 for 99 UI; 1/5 for IC), with measured detections ranging from 0.08-0.66 ng/L. Imidacloprid-urea was not 100 detected in IC tap samples and ranged from 0.22-0.29 ng/L at UI taps (2/2). Clothianidin, 101 imidacloprid, and thiamethoxam were also present in raw, treated, and tap samples with 102 concentrations ranging from 2.34–25.34 ng/L for clothianidin, 1.02–8.79 ng/L for imidacloprid, 103 and 0.24-5.99 ng/L for thiamethoxam. Notably, tap water concentrations for both UI and IC were similar to those we previously reported²⁸ (Table S.5). In contrast to our previous study, we 104 105 observed removal of clothianidin and imidacloprid between the source and treated UI samples. We 106 attribute removal to a powder activated carbon system that was added to UI for control of disinfection byproduct precursors after our initial study. This updated system is likely also 107 108 removing neonicotinoid parent compounds, which we previously reported were effectively removed via activated carbon.28 109

Although the concentrations of metabolites were substantially lower than their respective 110 parent compounds, select neonicotinoid metabolites are known to exhibit higher mammalian 111 toxicity, based on limited available data. Desnitro-imidacloprid has a substantially lower IC₅₀ value 112 than imidacloprid for vertebrates, indicating greater binding response (8.2 vs 2600 nM [1.7 vs 550 113 $\mu g/L$], respectively).³ The greater potential toxicity and frequent presence in these water samples 114 115 of neonicotinoid metabolites demonstrates the need to consider their fate and persistence in drinking water treatment systems (e.g., during chlorination and other treatment processes) and 116 their potential effects on human health. Indeed, neonicotinoids have been measured year-round¹⁰ 117 in streams of impacted watersheds, and our results demonstrate that consumers of drinking water 118 derived from vulnerable sources may be exposed to neonicotinoids and their metabolites.²⁸ 119



Figure 1: Clothianidin, imidacloprid, thiamethoxam, and two metabolites of imidacloprid (imidacloprid-urea and desnitro-imidacloprid) measured in raw and treated water from the University of Iowa and Iowa City water treatment plants (July 23 and 24, 2018, respectively). University of Iowa tap water was collected at two locations and Iowa City tap water was collected from three residences across Iowa City (n=2 and 3, respectively, July 17, 2018). Tap

concentrations are reported as averages (n=3, July 17, 2018), where (*) denotes non-detects while (†) denote samples present below the lower detection limit (LLD). LLD values (ng/L): clothianidin, 0.488; imidacloprid, 0.275; thiamethoxam, 0.081; desnitro-imidacloprid, 0.026; imidacloprid-urea, 0.057. Error bars represent the standard error including the variation between samples and in sample processing/analysis (associated with the composite enrichment, sample extraction, and analysis).

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Desnitro-imidacloprid and Imidacloprid-urea Reactivity with Chlorine. Desnitro-133 imidacloprid and imidacloprid-urea react relatively rapidly during chlorination (Figure 2). Second-134 order rate coefficients (±SE) for imidacloprid-urea (2.7±0.2 M⁻¹s⁻¹) and desnitro-imidacloprid 135 136 $(72\pm5 \text{ M}^{-1}\text{s}^{-1})$ chlorination were calculated from measured pseudo-first-order rate constants (Figure S1-S2) assuming a constant HOCl concentration during reaction ($k_2 = k_{obs}/[HOCl]$). At a typical 137 chlorine concentration for disinfection (i.e., 5 mg/L as Cl₂) and assuming a constant residual, half-138 139 lives for imidacloprid-urea and desnitro-imidacloprid would be ~ 1.0 h and ~ 2.4 min, respectively. As such, the metabolites of imidacloprid could be expected to degrade readily in a chlorine 140 141 contactor and during distribution.

Notably, the half-life of desnitro-imidacloprid is much shorter than those we previously 142 reported for clothianidin, imidacloprid, or thiamethoxam²⁸—on the order of minutes compared to 143 hours or days for other neonicotinoids. We hypothesize that tautomerization³⁴ within the guanidine 144 functionality of desnitro-imidacloprid (Figure 2, Scheme S2) contributes to its greater reactivity, 145 resulting in an amino tautomer that would be expected to rapidly chloraminate based on the high 146 reactivity of primary amines toward free chlorine.³⁵ It remains unclear why imidacloprid-urea is 147 faster reacting than clothianidin and imidacloprid. Secondary and tertiary amides, such as those in 148 imidacloprid-urea, are known to be several orders of magnitude less reactive toward hypochlorous 149 acid than imine and guanidine analogs.³⁶ We therefore attribute the lower reactivity of clothianidin 150

and imidacloprid relative to imidacloprid-urea to the well-established electron-withdrawing nature
 of the nitro group.³⁷

153 Using HR-MS/MS fragment analysis, we propose structures for byproducts observed during 154 chlorination of desnitro-imidacloprid and imidacloprid urea. Chlorination of desnitro-imidacloprid results in the formation of two major identifiable products (hereafter desnitro-IMI 245 and 155 156 desnitro-IMI 279), corresponding to the addition of either one or two chlorines (*i.e.*, the formation 157 of one dichloro- and one trichloro-transformation product, respectively). Analysis of HR-MS/MS 158 fragmentation patterns indicates chlorine addition occurring in the guanidine-containing portion 159 of the molecule rather than the chloro-pyridine moiety (Fig S19) most likely via N-Cl bond formation; however, the exact site cannot be determined and thus desnitro-IMI 245 is reported at 160 a Level 3 confidence.³² Consistent with the formation of reactive N-Cl compounds, addition of 161 excess sulfite to product mixtures after desnitro-imidacloprid chlorination resulted in the loss of 162 163 detectable products and a corresponding increase in desnitro-imidacloprid (Figure S3). Such 164 byproduct reversibility in the presence of a reducing agent is indicative of chloramine formation, as has been previously reported during chlorination of amine-containing pharmaceuticals.³⁸ 165 Notably, this instability of desnitro-imidacloprid chlorination products may help to explain our 166 167 detection of desnitro-imidacloprid in finished tap water (Figure 1) despite its very high reactivity 168 toward free chlorine; decomposition of reactive byproducts could result in its regeneration during dechlorination with a reductant or via incidental reactions that occur within the distribution system. 169

We propose that desnitro-IMI 245 forms via chloramination of the amino tautomer of desnitroimidacloprid (Figure 2, Scheme S2), which we expect to preferentially chlorinate prior to the corresponding imino tautomer based on established trends in the chlorination of structurally analogous N-containing compounds.^{36,39} At higher chlorine concentrations or contact times, we

further hypothesize that sequential chlorination of desnitro-IMI-245 occurs through a chlorimino derivative, where the added chlorine stabilizes the imino tautomer akin to the electron-withdrawing nitro-group in imidacloprid. Although speculative, the secondary amine moiety in the chlorimino tautomer would again be expected to exhibit greater reactivity toward chlorine than the corresponding imine moiety in the chloramino tautomer.

179 Chlorination of imidacloprid-urea yielded one major identifiable product (hereafter IMI-urea
180 246). This corresponds to the addition of chlorine to the imidacloprid-urea structure. Once again,
181 HR-MS/MS fragment analysis is most consistent with chlorination occurring at the secondary
182 amide (Figure 2; Figures S39-S40).

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Figure 2: Chlorination of (left) desnitro-imidacloprid and (right) imidacloprid-urea to form 185 chlorinated products desnitro-IMI 245, desnitro-IMI 279 and IMI-urea 246. Chlorination kinetics 186 of desnitro-imidacloprid to desnitro-IMI 245 and imidacloprid urea to IMI-urea 246 are shown. 187 Peak area shown is the HPLC-DAD response $\lambda = 260$ nm for imidacloprid-urea; 273 nm for 188 189 desnitro-IMI; relative values are shown because no authentic standards of chlorinated products are available. Initial concentration conditions (molar ratios shown in figures): desnitro-imidacloprid= 190 10µM, 1 mg/L HOCl as Cl₂; imidacloprid urea= 5µM, 1 mg/L HOCl as Cl₂. Full kinetics data, 191 conditions in Figures S1, S2. 192

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194 Hydrolysis Products of Thiamethoxam and Reactivity with Chlorine. The alkaline hydrolysis of thiamethoxam (at pH 10; relevant to lime softening) results in two products (hereafter THX-H 195 196 248 and THX-H 237), both of which have been previously identified with proposed pathways.^{5,15,20,40} Imines are known to easily hydrolyze in water to yield ketones,^{41,42} and the 197 198 electron-withdrawing -NO₂ substituent makes the carbon in the guanidine portion of thiamethoxam more electrophilic, thus inviting hydroxide attack under alkaline conditions.²⁰ THX-H 248 is 199 200 formed through the simple hydrolysis of the nitro-imine group into a ketone.²⁰ THX-H 237 was reported by Maienfisch⁵ and corresponds to a ring opening with hydroxide attack at the imine 201 carbon. 202

Upon addition of chlorine, THX-H 237 is reactive, while THX-H 248 is recalcitrant over the timescales / conditions investigated (Figure 3). We attribute the greater reactivity of THX-H 237 toward chlorine to the presence of its two secondary amides. The second-order rate coefficient $(\pm SE)$ for the reaction of free chlorine with THX-H 237 (0.67±0.02 M⁻¹s⁻¹) was calculated from the measured pseudo-first-order rate constant (Figure S4). Assuming a constant chlorine residual (5 mg/L Cl₂), the half-life of THX-H 237 would be 4.8 h.

THX-H 237 reacts with chlorine to produce a single species hereafter referred to as CLO-THX H 270 (see also Table 1). We propose that chlorine addition occurs at the secondary amide group

without the electron-withdrawing nitro substituent (Figure 3). Our MS/MS fragmentation results
reveal a corresponding chlorinated fragment to support this proposed structure (Figure S9-S10).
We anticipate that THX-H 237 will react to generate CLO-THX-H 270 at time-scales relevant to
disinfection and distribution in systems that also employ chemical (e.g., lime-soda) softening
earlier in the treatment process train.

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Figure 3: Chlorination of the hydrolysis products of thiamethoxam (THX-H 237 and THX-H 248) to form novel chlorinated product CLO-THX-H 270. Chlorination kinetics (represented by HPLC-DAD peak area $\lambda = 260$ nm; authentic standards unavailable; 50 mg/L Cl₂) of thiamethoxam hydrolysis product THX-H 237 to CLO-THX 270 is shown (THX-H 248 was unreactive) at pH 10. The structure of CLO-THX 270 (the same as generated through chlorination of clothianidin) is presented as shown to be consistent with Table 1; chlorination occurs at either amine farther from the nitro-group as determined by HR-MS/MS fragmentation (Figures S9, S10, S34, S35).

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Products of Imidacloprid and Clothianidin Chlorination. We previously reported timescales for the reaction of imidacloprid and clothianidin with chlorine.²⁸ Herein, we propose structures using the Schymanski framework³² to communicate confidence of novel products discovery for the products of these reactions (Table 1) based on HR-MS/MS fragment analysis of these product
 mixtures (Table S.7 describes compounds prior to chlorination).

231 Chlorination of clothianidin results in three major products. Two products have with the same 232 mass (hereafter CLO-239a and CLO-239b) but different retention times, while the third has an exact mass [M+H]⁺ of 270.9442. The latter product appears identical to the product formed during 233 234 chlorination of thiamethoxam hydrolysis products, and is thus also referred to as CLO-THX-H 235 270. Clothianidin is a known product of thiamethoxam degradation through multiple reported biologically-mediated pathways^{43,44} (e.g., in insects, mammals, plants, and soil) where the two 236 compounds share common metabolites;^{25,45} however, abiotic and biological pathways may 237 generate different products. CLO 239a and CLO 239b correlate to loss of the nitro group, 238 239 formation of the ketone (C=O), and chlorination of a remaining secondary amide. We suspect these reactions occur in a step-wise fashion and involve both oxidation with chlorine and hydrolysis 240 (e.g., imine hydrolysis to a ketone) reactions, potentially involving intermediates we were unable 241 242 to identify. The exact location of the chlorine on two of the clothianidin products (CLO 239a, CLO 239b) could not be confirmed with certainty because MS/MS fragmentation did not yield the 243 chlorinated component (Figure S11-S14; Level 3 confidence). Nevertheless, chlorination is most 244 245 likely to occur at either of the secondary amides because HR-MS/MS fragment analysis indicated that the chlorothiazole component was not further chlorinated (Figure S8-S14). Fragmentation 246 247 analysis of CLO-THX-H 270 generated either with clothianidin or thiamethoxan as the parent 248 compounds suggests that chlorination occurs at the nitrogen farther from the nitro group because a chlorinated fragment consistent with this structure was present (Figures S9, S10; S34, S35; Level 249 2b confidence). 250

251	Chlorination of imidacloprid forms three major transformation products (hereafter: IMI-urea
252	246, IMI 290, and IMI 341). Product IMI-urea-246 is chlorinated imidacloprid-urea, which we
253	previously identified in our independent analysis of products generated from the chlorination of
254	an imidacloprid-urea standard (described above). IMI 290 is chlorinated imidacloprid (without
255	loss of the nitro group), with chlorination most likely occurring at the secondary nitrogen in its
256	guanidine moiety. One product, IMI 341, could only be confirmed to level 5 confidence, ³² thus no
257	structure is proposed.

- Table 1: Transformation products of clothianidin, imidacloprid, desnitro-imidacloprid, 258
- imidacloprid-urea, and thiamethoxam. 259

	Neonicotinoid Chlorination and Hydrolysis Transformation Products						Fragment lons	
Parent Compound	Product Name	Proposed Structure	Proposed Formula	Schymanski [‡] Confidence Level	RT (min)	Accurate Mass [M+H] ⁺	Accurate mass (m/z)	Proposed Molecular Formula
Clothianidin	CLO 239 a	$\begin{array}{c} CI \xrightarrow{S} \\ N \xrightarrow{O} \\ Or \\ CI \xrightarrow{S} \\ N \xrightarrow{O} \\ CI \xrightarrow{S} \\ CI \xrightarrow{V} \\ CI \xrightarrow{V} \\ CI \xrightarrow{V} \\ CI \xrightarrow{V} \\ H \end{array} \xrightarrow{CH_3} $	C ₆ H ₇ Cl₂N₃OS	Level 3	16.1	239.9792	168.0261 174.9774 204.0124 119.9693 86.0095	$\begin{array}{c} C_6H6N_3OS\\ C_5H_4CIN_2OS\\ C_6H_7CIN_3OS\\ C_3H_2CINS\\ C_3H_3NS \end{array}$
Clothianidin	CLO 239 b	$\begin{array}{c} CI \xrightarrow{S} \\ N \xrightarrow{O} \\ r \\ CI \xrightarrow{S} \\ N \xrightarrow{O} \\ CI \xrightarrow{S} \\ N \xrightarrow{O} \\ CI \xrightarrow{S} \\ CI \xrightarrow{O} \\ H \\ CI \xrightarrow{H} \\ H \\ CI \xrightarrow{H} \\ H \\ H \end{array}$	C ₆ H ₇ Cl₂N₃OS	Level 3	16.4	239.9798	174.9771 146.982 131.9711 168.0261 119.9788	$\begin{tabular}{lllllllllllllllllllllllllllllllllll$
Clothianidin, THX-H 237	CLO-THX-H 270	$CI \xrightarrow{S} N \xrightarrow{O} N^{-NO_2}$	$C_5H_4Cl_2N_4O_3S$	Level 2b	9.2	270.9442	181.9439 146.9768 132.9717 118.9552	$\begin{array}{c} C_4H_3Cl_2N_2S\\ C_4H_5CIN_2S\\ C_4H_4CINS\\ C_3HCINS \end{array}$
Imidacloprid, Imidacloprid- urea	IMI-urea 246		C ₉ H ₉ Cl ₂ N ₃ O	Level 2b	15.9	246.0222	211.0487 155.0348 141.0206 126.0097	$\begin{array}{c} C_9H_{10}CIN_3O\\ C_7H_8CIN_2\\ C_6H_6CIN_2\\ C_6H_5CIN \end{array}$
Imidacloprid	IMI 341	Unknown	Ambiguous	Level 5	16.6	341.9938	218.0239 155.0367 126.0104	Unknown Unknown Unknown
Imidacloprid	IMI 290		$C_9H_9Cl_2N_5O_2$	Level 2b	16.9	290.0222	246.0217 209.0617 173.0839 126.0123	$\begin{array}{c} C_{9}H_{9}CI_{2}N_{3}O\\ C_{9}H_{10}CIN_{4}\\ C_{9}H_{10}N_{4}\\ C_{6}H_{5}CIN \end{array}$
Thiamethoxam	ТНХ-Н 237		C ₅ H ₅ ClN ₄ O ₃ S	Level 2b	11.8	236.9838	174.9724 147.9772 97.0388	$C_{5}H_{4}CIN_{2}OS$ $C_{4}H_{5}CIN_{2}S$ $C_{4}H3NS$
Thiamethoxam	ТНХ-Н 248		C ₈ H ₁₀ ClN ₃ O ₂ S	Level 2a	11.2	248.0248	174.9718 98.0048 131.9665	$\frac{C_5H_4CIN_2OS}{C_4H_4NS}$ C_4H_3CINS
Desnitro- imidacloprid	desnitro-IMI 245		$C_9H_{10}Cl_2N_4$	Level 3	14.7	245.0377	209.0622 173.0848 211.0766 83.0588 132.0353 126.0133	$\begin{array}{c} C_{9}H_{10}CIN_{4}\\ C_{9}H_{10}N_{4}\\ C_{9}H_{11}CIN_{4}\\ C_{3}H_{5}N_{3}\\ C_{4}H_{7}CIN_{3}\\ C_{6}H_{5}CIN \end{array}$
Desnitro- imidacloprid	desnitro-IMI 279		$C_9H_9CI_3N_4$	Level 2b	18.5	279.0004	209.0506 173.0848 126.0130	$C_{9}H_{10}CIN_{4}$ $C_{9}H_{10}N_{4}$ $C_{6}H_{5}CIN$

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[‡]The confidence level and structure of each product is characterized according to the Schymanski et al. 2014 framework for identifying small molecules via high resolution mass spectrometry.³²All samples were 262 analyzed in in ESI positive mode (i.e., ion [M+H]⁺ =compound exact mass+H). High-resolution 263 fragmentation patterns are presented in Figures S6-S40. 264

Environmental Implications. This is the first known study to report neonicotinoid metabolites in 265 drinking water, and builds upon our prior research²⁸ and a subsequent publication from Canada⁴⁶ 266 demonstrating neonicotinoids in drinking water. We also show that neonicotinoids and their known 267 metabolites can form transformation products during disinfection and/or lime softening 268 (hydrolysis at elevated pH) at timescales relevant to water treatment / distribution. The mammalian 269 270 toxicity of transformation products formed during water treatment processes remains unknown. It is possible that chlorination of neonicotinoids and their metabolites will impact receptor binding 271 interactions and alter their bioactivity relative to that of the parent neonicotinoids or known 272 273 metabolites, a scenario that requires further investigation. Several transformation products identified (CLO 239a, CLO 239b, CLO-THX-H 270, IMI 246, THX-H 248, DN-IMI 245 and DN-274 IMI 279) appear to lose the nitro-group through chlorination or hydrolysis, and/or gain one or more 275 276 chlorines—both characteristics that may increase mammalian toxicity.^{3,4,23,29,31,47} Additional studies are needed to better assess temporal and spatial trends in metabolite occurrence / toxicity 277 278 of chlorinated DBPs formed during drinking water treatment (including synthesized standards), especially in waters impacted by parent neonicotinoid insecticides. 279

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SUPPORTING INFORMATION. Additional method details, statistical analysis, quality
 assurance / control, additional detailed data / results / analysis in figures and tables.

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